

# Emission of CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O during Composting of Two-Phase Olive Mill Wastes: A Four-Year Study

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## ABSTRACT

A four-year monitoring program was performed to quantify the emission of the three most important greenhouse gases (GHG),  $CO_2$ ,  $CH_4$  and  $N_2O$ , emitted during composting of two-phase olive mill wastes (TPOMW). A composting pile was prepared with typical agroindustrial wastes generated from olive oil mills: TPOMW and olive tree prunings, used as bulking agent, enriched with urea as N source. The study was performed during four consecutive years in order to obtain a wide set of experimental data to evaluate the effect of the variability of the raw material characteristics and the performance of the composting plant on GHG emissions. Biogenic  $CO_2$  was the main gas generated during TPOMW composting. The range of  $CO_2$  emission exhibited low interannual variation regardless the operational conditions at the composting plant. The improvement of the composting operations (better control of pile moisture and improved aeration) reduced the amount of  $CH_4$  and  $N_2O$  emitted from the pile by reducing the formation of anaerobic spots within the pile. The control of the  $CH_4$  and  $N_2O$  emissions reduced the contribution of TPOMW composting to climate change, expressed as kg of  $CO_2$ equivalents.

Keywords: composting, greenhouse gas, methane, nitrous oxide, manure, olive mill wastes

## INTRODUCTION

The interest on the evaluation of the environmental impact of waste management on greenhouse gas emissions (GHG) is growing over the last decades. The most important greenhouse gases generated during organic waste composting are  $CO_2$ ,  $CH_4$  and  $N_2O$ .  $CO_2$  is the most abundant gas generated from organic matter degradation, however, it is not considered for the evaluation of the global warming potential (GWP) since it has a biogenic origin (IPCC 2006). For this reason  $CH_4$  and  $N_2O$  have attracted the attention of researchers in recent years since their global warming potential is about 25 and 298 times more potent than carbon dioxide, respectively (over a 100 year period).

The emission of these GHG has been recently reviewed for different composting operations (Brown et al. 2008) and different feedstocks such as manures (Kebreab et al. 2006) and municipal solid wastes (Lou and Nair 2009). CH<sub>4</sub> is generated in anaerobic spots within the composting matrix even in well managed piles. The high microbial activity characteristic at early stages of the process leads to the decline in O<sub>2</sub> concentration favouring the formation of anaerobic spots where CH<sub>4</sub> can be generated. In the case of N<sub>2</sub>O, its generation is commonly associated to nitrification and denitrification process taking place either at the beginning or after the bioxidative phase of composting, when pile temperatures are below 40°C. The characteristics of the raw materials and the composting operations can also affect the generation and diffusion of these gases to the atmosphere (excessive moisture content, poor aeration, physical compaction of the composting materials, etc). For this reason it is important to study the GHG emission for different organic materials in order to understand the mechanisms involved for each particular feedstock and different operational conditions.

Composting of two-phase olive mill wastes (TPOMW), the main by-product generated in olive oil processing indus-

try, represents and important waste management strategy in Mediterranean areas for the recycling of this particular waste (Roig et al. 2006). TPOMW are usually treated by windrowing in open air composting pads located close to the olive mills. Despite the increasing number of TPOMW composting plants in operation in Mediterranean areas there is only little information on the GHG emissions from such facilities and the effect of chemical composition of TPOMW on these emissions (Manios et al. 2007; Sánchez-Monedero et al. 2010). TPOMW is a semisolid waste characterised by a slightly acidic pH, a high organic matter content (mainly lignocellulosic), a relatively low amount of available N and the presence of phenols and fats (Roig *et al.* 2006). The large moisture content and the presence of fats in TPOMW favour the compaction of these wastes and prevent the correct aeration of the pile, increasing CH4 generation. The slow degradation rate of TPOMW and the limited amount of mineral N reduces the generation of both gases during composting but the addition of different bulking agents and N sources can markedly modify the emission rates (Sánchez-Monedero et al. 2010)

The aim of this study was to quantify the emissions of  $CO_2$ ,  $CH_4$  and  $N_2O$  during TPOMW composting during four consecutive years in order to evaluate the effect of the variability in the chemical composition or raw wastes and operational conditions on the emission of the most relevant GHG. The interannual evaluation would also allow to asses the contribution of TPOMW composting operations to climate change by calculating the global warming potential for each year as the sum of the total  $CO_2$ ,  $CH_4$  and  $N_2O$  emissions (expressed as kg of  $CO_2$  equivalents).

## MATERIALS AND METHODS

## **Compost elaboration**

Composts were prepared during four consecutive olive oil seasons

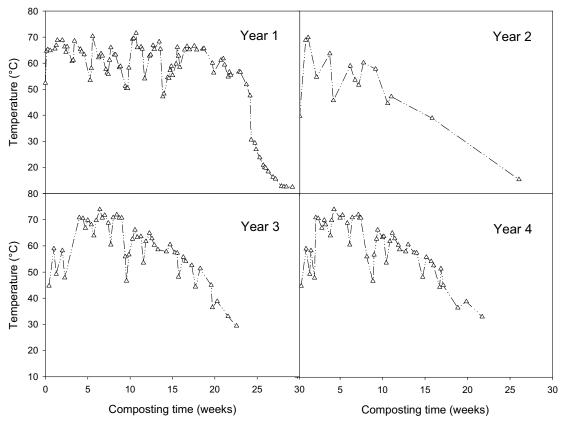


Fig. 1 Temperature evolution of the composting mixture in different production years.

in a pilot-scale composting plant treating the organic wastes obtained from the experimental farm "Venta del Llano" (IFAPA), located in Jaén (Spain). The same composting mixture was prepared every year by mixing fresh two-phase olive mill waste (TPOMW), olive tree prunings and urea. The dry weight proportion of the starting mixture was: TPOMW (30%) + olive tree pruning (70%) + urea (9 kg/tonne). This composting mixture represents one of the most typical compositions being composted in olive oil production areas since it combines the most abundant organic wastes generated in olive oil mills and olive orchards. This mixture was enriched with urea to achieve an initial C/N ratio in the range within 25-30.

About 5000 kg of this mixture was composted by windrowing in a trapezoidal pile (5 m length, 2 m width on the base and 1 m height) as already described for similar composting mixtures (Serramiá *et al.* 2010). The pile was turned every two weeks during the bio-oxidative phase (approximately three months), and the mixture was then allowed to mature over a period of two months. The moisture of the pile was controlled every two weeks and adjusted by adding the necessary amount of water to keep the levels close to 40% during the first two years and between 30 and 40% in years 3 and 4. Pile temperature was measured daily during the bioxidative phase of composting and then every week until the end of the process.

## Gas sampling and analysis

A static closed chamber technique was used to measure  $CO_2$ ,  $CH_4$ and  $N_2O$  fluxes from the surface of the composting piles (Sommer *et al.* 2004). Gas samples were taken once a week (during the biooxidative phase of composting) and approximately every two weeks (during maturation) from five aluminium closed chambers (volume: 0.010 m<sup>3</sup>, area: 0.038 m<sup>2</sup>) placed at different locations on the top of the pile. The chamber were pushed 5 cm into the composting pile and air samples were taken within the headspace at 0, 15 and 30 min using disposable syringes. Air samples were transferred to evacuated 5 ml glass vials (Vacutainer<sup>®</sup>) fitted with butyl rubber stoppers and analysed within 24 hours by gas chromatography (VARIAN CP-4900 Micro-GC). Emission fluxes were calculated by fitting the experimental data to a second-order polynomial equation (gas concentration vs time). The flux at time 0 was calculated by taking derivates of the second order polynomial (Hao et al. 2001).

The estimation of the global warming potential (GWP) of the composting operation was obtained by multiplying the cumulative mass of  $CH_4$  and  $N_2O$  emitted per year by 25 and 298, respectively, in order to obtain the total amount of equivalent  $CO_2$  (IPCC 2006).

#### Experimental design and statistical analyses

A composting pile was prepared during four consecutive years with the same initial composition (same proportion of organic wastes in the starting mixtures) in order to study the interannual evolution of the process. A GHG monitoring program, consisting of 18 to 20 sampling events, was performed every year during the composting process, as previously described. Gas fluxes were expressed as grams of C or N per unit of area ( $m^2$ ) and per time (day) and represented the mean and standard error of 5 replicates (n=5).

### **RESULTS AND DISCUSSION**

#### Performance of the composting process

The general performance of the composting process was evaluated by the evolution of the temperature of the composting piles (Fig. 1). The four composting piles exhibited the typical pattern of olive mill waste composting, characterized by a long duration of the process, up to 35-40 weeks, as a consequence of the lignocellulosic composition of the material and the low amount of available N (Cayuela et al. 2006; Alburquerque et al. 2009). The piles were characterized by long bioxidative phases (around 15 weeks) with high temperatures above 50°C and a maturation phase where pile temperature slowly decreased. The different temperature evolution observed during the four years of the experiment may reflect the variation in the chemical and physicochemical characteristics of the starting materials and also the performance of the composting plant. Since only minor differences in chemical composition of TPOMW and olive tree prunings were expected within olive oil seasons, the composting plant operation can play a key role on the general performance of the process. Composting piles prepared

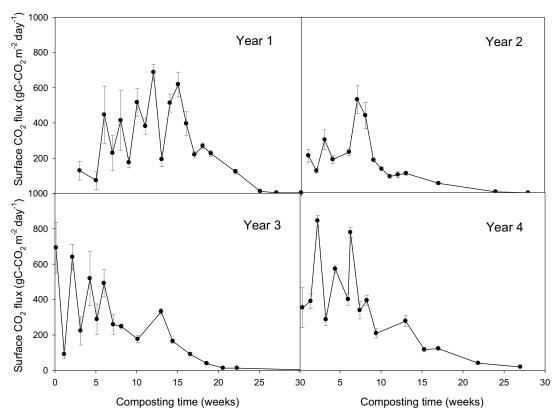


Fig. 2 Surface CO<sub>2</sub> fluxes emitted from the composting pile during different production years. Bars represent the standard error (n = 5).

in years 3 and 4 showed very similar patterns reflecting the experience reached by the composting plant operators in optimizing the composting process (pile size, moisture control, turning skills, etc.). Sánchez-Monedero *et al.* (2010) reported the variation of the typical stabilisation indices during the first year of composting operation.

#### CO<sub>2</sub>, CH<sub>4</sub> and N<sub>2</sub>O production during composting

CO<sub>2</sub> was the main gas generated during composting as a consequence of microbial degradation of organic matter. Thus, the emission of CO<sub>2</sub> was closely related to the activity of the composting pile, and both surface CO<sub>2</sub> flux (Fig. 2) and temperature (Fig. 1) showed parallel trends in the four composting piles. The largest CO<sub>2</sub> fluxes were registered within early stages of the process, during the bioxidative phase whereas a steady decline of CO<sub>2</sub> emission was observed during the maturation phase, corresponding to a decrease in microbial activity. Biweekly turnings during the bioxidative phase originated peaks of CO<sub>2</sub> emissions as a consequence of reactivation of the composting process (homogenisation and water addition). Surface  $CO_2$  fluxes ranged between 200 and 800 gC-CO<sub>2</sub> m<sup>-2</sup> day<sup>-1</sup> in the first half of the process and they were reduced to less than 10  $gC-CO_2 m^{-2} day^{-1}$  by the end of maturation. The surface gas fluxes showed a homogeneous distribution along the upper part of the pile surface and showed a relatively low coefficient of variation.

The range of surface  $CO_2$  fluxes only showed little interannual variation reflecting that gas emission was not greatly affected by the origin of the starting materials or the operating conditions at the composting plant. During the first two years of operation, a lag phase was observed in the  $CO_2$  production which was probably caused by the performance of the composting plant and the moisture of the piles (higher during the fist two years of the study). After these initial years, the composting pile showed very similar  $CO_2$ emission patterns as a consequence of the optimisation of the composting operations (turning and moisture) and the training of plant operators.

Despite the prevalent aerobic conditions in composting

piles, CH<sub>4</sub> can be generated at early stages of the process when the available O<sub>2</sub> is partially depleted as a consequence of the large microbial activity and high temperatures. Surface CH<sub>4</sub> fluxes were generally lower than 30 gC-CH<sub>4</sub> m<sup>-2</sup> day<sup>-1</sup> (**Fig. 3**) during composting, which are below the common surface CH<sub>4</sub> fluxes registered during composting of manures such as beef feedlot and liquid swine, from 20 to 1250 gC-CH<sub>4</sub> m<sup>-2</sup> day<sup>-1</sup> (Kebreab *et al.* 2006). During the first two years of operation, large peaks of CH<sub>4</sub> production, up to 170 gC-CH<sub>4</sub> m<sup>-2</sup> day<sup>-1</sup>, were registered during the active phase of composting. Contrary to CO<sub>2</sub> emissions, surface CH<sub>4</sub> fluxes showed a large coefficient of variation along the pile surface, reflecting the presence of small anaerobic spots within the pile, rather than a uniform distribution of CH<sub>4</sub> production.

The large interannual variation exhibited by the surface  $CH_4$  fluxes suggests that the emission of  $CH_4$  is likely to be affected by the operational conditions and experience of site workers. In fact, the emission of  $CH_4$  was markedly reduced in the last two years of operation, where the peaks of production were significantly reduced by an improved performance of the process. In the last two years of the experience the moisture levels were kept lower than in previous years 1 and 2 to avoid the compaction of the material. Turnings were also improved in the last to years to allow a better aeration and homogenisation of the composting pile.

The emission of N<sub>2</sub>O was generally low (below 200 mgN-N<sub>2</sub>O m<sup>2</sup> day<sup>-1</sup>) except in year 2 where two peaks up to 1200 mgN-N<sub>2</sub>O m<sup>-2</sup> day<sup>-1</sup> were registered at the end of the bioxidative phase (**Fig. 4**), when pile temperature started to decline. Sanchez-Monedero *et al.* (2010) found that N<sub>2</sub>O emission was highly dependent on the N source used for olive mill waste composting and they observed that the use of urea could significantly increase the generation of N<sub>2</sub>O due to the large availability of mineral N (ammonium and nitrate) in the mixture. The peak observed in year 2 may be related to the limited O<sub>2</sub> availability in the composting mixture, which is a key element controlling the N<sub>2</sub>O generation during composting (Czepiel *et al.* 1996), and also to the decrease in pile temperature below 50°C after the 10<sup>th</sup> week (**Fig. 1**), which allowed the activity of nitrifying bacteria. In

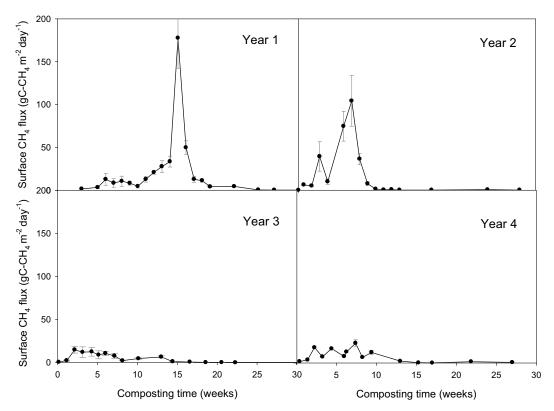


Fig. 3 Surface CH<sub>4</sub> fluxes emitted from the composting pile during different production years. Bars represent the standard error (n = 5).

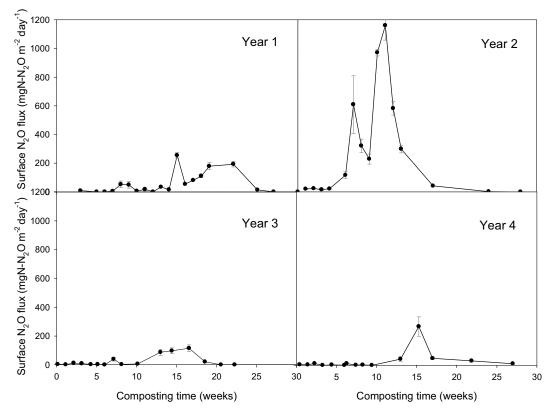


Fig. 4 Surface N<sub>2</sub>O fluxes emitted from the composting pile during different production years. Bars represent the standard error (n = 5).

the case of TPOMW composting, Sánchez-Monedero *et al.* (2010) reported that nitrification was thought to be the main source of  $N_2O$  emission which was affected by the presence of mineral N and  $O_2$  availability in the pile.

Surface  $N_2O$  fluxes exhibited a homogenised distribution along the pile as observed in the low coefficient of variation of the measurements. However, the peaks observed in year 2 were characterised by large standard deviation, as already observed in CH<sub>4</sub> emission. The large spatial variability of  $N_2O$  emissions can be due to the presence of anaerobic spots within the pile, with limited  $O_2$  availability, which favours the formation of  $N_2O$  by nitrifying bacteria (Béline *et al.* 1999).

## Evaluation of GWP of olive mill waste composting

Global warming potential (GWP) was calculated as the sum of total kg of  $CO_2$ ,  $CH_4$  and  $N_2O$  (expressed as kg of  $CO_2$ 

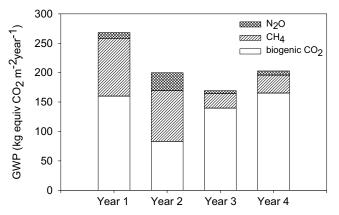


Fig. 5 Global warming potential (GWP) for each year calculated as the sum of total kg of  $CO_2$ ,  $CH_4$  and  $N_2O$  (expressed as kg of  $CO_2$  equivalents) per compost surface unit.

equivalents) per compost surface unit in order to compare the contribution to climate change of each greenhouse gas emitted during composting (**Fig. 5**). During the four campaigns CO<sub>2</sub> represented the gas with the highest contribution, however, according to Smith *et al.* (2001) for biodegradable materials which carbon have been recently absorbed from the atmosphere by photosynthesis: "... *there is no net* global warming impact, and no global warming potential is associated with the CO<sub>2</sub> emission, .... These emissions are reported as 'short term CO<sub>2</sub>' or 'biogenic CO<sub>2</sub>' and are given a global warming potential of zero."

The second most important contribution was  $CH_4$ , which GWP was substantially high the first two years. Nevertheless,  $CH_4$  emissions were successfully reduced the last two years mainly as a consequence of moisture optimization and the acquired experience of the operators at the composting facility for improved aeration.

 $N_2O$  emissions were more difficult to control since  $N_2O$  can be produced during both nitrification (aerobic) and denitrification (anaerobic) processes and it seems to be mostly influenced by the original materials used for composting. However, its share to total GWP was always the lowest.

## CONCLUSIONS

The levels of  $CH_4$  and  $N_2O$  emissions generated during TPOMW composting were relatively low compared to other organic waste such as manures and municipal solid wastes. A proper management of composting operations, in terms of improved aeration and moisture control, can significantly reduce the contribution of TPOMW composting to global change, as far as GHG are concerned.

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